

# Short-term exposure assessment to particulate matter and total volatile organic compounds in indoor air during cooking Ethiopian sauces (*Wot*) using electricity, kerosene and charcoal fuels

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## Abstract

A total of 45 households were selected to participate in a study to assess indoor exposure to particulate matter (PM) and total volatile organic compounds (TVOC) while cooking a widely consumed Ethiopian traditional stew sauce (*Wot*, in Amharic). The geometric mean (GOM) concentrations of PM<sub>1</sub>, PM<sub>2.5</sub> and PM<sub>10</sub> when electricity, kerosene and charcoal fuel sources were used during the wet season ranged 11.4–151 µg.m<sup>-3</sup>, whereas the GOM concentrations during the dry season ranged 7.7–222 µg.m<sup>-3</sup>. The GOM of TVOC during the wet and dry seasons using electricity, kerosene and charcoal fuel ranged 350–812 µg m<sup>-3</sup>. The health risks associated with exposure to PM<sub>2.5</sub>, PM<sub>10</sub> and total suspended particulate matter (TSP), either as a single pollutant or cumulative based on the hazard quotient (HQ) or hazard index (HI) calculation by U.S. Environmental Protection Agency, indicate that exposure to PM<sub>2.5</sub>, PM<sub>10</sub> or TSP while cooking *Wot* using electricity, kerosene and charcoal fuels may not cause health problems to a healthy adult. However, long-term exposure to high levels of PM<sub>10</sub> emissions when charcoal fuel is used is considered to be unsafe. The exposure to all the emitted pollutants during the use of any of the three fuels sources may also pose a human health impact to individuals near the sources.

## Keywords

Charcoal, Kerosene, Electricity, Indoor air pollution, Particulate, VOC, *Wot*, Ethiopia

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## Introduction

More than 40% (>3 billion people) of the world's population rely on unprocessed solid biomass fuels (wood, charcoal, agricultural residues and animal dung) for their daily household energy needs.<sup>1,2</sup> Such types of fuel use are much higher in developing countries than in developed countries. Studies in Sub-Saharan Africa have shown that 90–95% of domestic energy (for cooking, lighting and heating) depends on the biomass fuels, with cooking having a significant share. In Ethiopia,

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more than 90% of the population use biomass fuels for cooking, heating and lighting, of which 99% is derived from firewood, charcoal, crop residue and leaves.<sup>3,4</sup>

Exposure to high levels of air pollution has been consistently associated with different types of diseases, such as the risk of respiratory tract infections, exacerbations of inflammatory lung conditions, cardiac events, stroke, eye disease, ischaemic stroke, tuberculosis and cancer.<sup>5-14</sup> However, the type of diseases caused by air pollution depends on the pollutant concentration in the air, the type of pollutants, the composition of pollutants (which is affected by its sources) and the length of time that an individual inhales the pollutants. Thus, chronic exposures to low levels of pollution might result in an equivalent dosage to a person who undergoes acute exposures to a high level of air pollution. For example, according to WHO, women exposed to biomass smoke for 2 to 4 h during cooking can inhale total suspended particulates (TSP) and benzo ( $\alpha$ )pyrene in an amount equivalent to smoking 40 cigarettes in 24 h.<sup>15</sup>

Generally, among the substances in polluted air, particulate matter (PM) with different aerodynamic diameters and total volatile organic compounds (TVOC) are major components and they are a health concern in the world nowadays.<sup>16,17</sup> Carbon-based compounds that have high vapour pressures and significantly vaporize and enter the atmosphere are called volatile organic compounds (VOCs).<sup>18</sup> TVOC and PM with different aerodynamic diameters released during the combustion of biomass fuel and traffic emissions are a significant problem for the developing world in general and Ethiopia in particular.<sup>19,20</sup> The comprehensive exposure assessments of PM with different aerodynamic diameters and TVOC are used for estimating the health burden of people, for assessing the contribution levels of various sources and for developing strategies for management.

Thus, several studies have been carried out around the globe on the exposure assessment of different pollutants in different scenarios. The exposure assessment to indoor air PM and TVOC during different activities is very important for estimating the contribution of each activity to the total daily exposure. Moreover, recently, the studies also focused on identification of high pollutant-emitting activity and personal exposure assessment, which are used to design and take the intervention at the individual level.<sup>21-25</sup> Likewise, in Ethiopia, very few cross-sectional and pilot studies have been conducted on exposure assessment for both indoor and outdoor air pollution. Hence, different researchers have reported concentrations of CO, NO<sub>2</sub>, PM<sub>10</sub>, PM<sub>2.5</sub> and TSP in indoor and outdoor air at different parts of the country (Gonder, Butajira area, Shebidino wereda, rural area of Tigray,

Kebribeyah, rural area of Jima, rural area of Wellega and Gimbie and Addis Ababa). Most of the levels of indoor air pollution exceeded United States Environmental Protection Agency (US EPA) and WHO guidelines.<sup>26,27</sup> This might be due to living in crowded and poorly ventilated housing, and to limited access to separate cooking and living areas in rural and urban homes, which is a common feature of living conditions in Ethiopia.<sup>2,6,15,28-32</sup> In general, the air pollution and its health impact were studied by the measurement of air pollutants at a fixed place, which does not predict accurate exposure assessment for pollutants. Thus, using personal exposure data at different microenvironments rather than fixed monitoring data is the better method in exposure assessment and in identifying the role of each microenvironment to personal exposure.<sup>33-35</sup> However, almost all exposure assessment studies for PM in Ethiopia have been based on 24-h measurements at fixed sites for indoor and outdoor air, which might not show the real exposure of a person. This is because the person does not stay in the home or outdoor for 24 h. Although acute exposure to a high level of PM has been shown to cause health problems, studies conducted in Ethiopia lack information on quantification and assessment of short-term exposure (from one to several hours) to PM from different activities.<sup>20,36,37</sup> There are a limited number of rigorous studies conducted in Ethiopian households that quantify indoor microenvironment TVOC levels from everyday activities and that assess the health impacts of short-term exposures.

Consequently, real-time personal monitoring coupled with individual activity data are required for assessing the exposure risk of population groups and the level and frequency of exposure. This type of measurement should be carried out at the microenvironment level, especially where high exposure occurs.<sup>38</sup> Nevertheless, none of the Ethiopian studies conducted so far measured at specific microenvironments. Among the cooking activities using biofuels in Ethiopia, cooking of the most widely consumed sauce called *Wot* is the major activity. People cooking such type of stew have used different kinds of fuels, namely electricity, kerosene and charcoal, which are widely used fuels for cooking *Wot* in urban areas. Among these fuels, charcoal is the most widely used fuel.<sup>39</sup> Also, the level of PMs and TVOC emitted from different fuel types in preparing *Wot* might be different. Hence, measurement of such pollutants is vital for selecting the better fuel type and estimating the health impacts of the *Wot* cooker. However, to the best of our knowledge, no studies have been carried out for quantification of PM and TVOC while cooking *Wot* using charcoal,

kerosene or electricity in households at Addis Ababa or anywhere in Ethiopia.

Therefore, the present study focuses on (a) assessing short-term exposure to PMs and TVOC during the cooking of *Wot* in 45 individual households using different biomass fuel types, (b) comparing the level of emissions during the cooking of *Wot* using charcoal, kerosene and electricity and (c) estimating the health risk to the person cooking due to lifetime exposure to PM<sub>2.5</sub>, PM<sub>10</sub> and TSP during the cooking of *Wot*.

## Material and methods

### Description of study area

Ethiopia is the second most populous country in Africa with a population of 102.4 million in 2016, while the population is predicted to reach 108.1 million by the end of 2018. The metropolitan city of Addis Ababa has a population exceeding three million and has been growing at a rate of 2.1% from 1994 to 2010.<sup>19</sup> The city is situated at the centre of the country at an altitude between 2200 and 2800 m above the sea level, between latitude 9.0300°N and longitude 38.7400°E. Average minimum and maximum annual temperatures range from 9.53 to 23.2°C.<sup>40</sup>

Since the level of outdoor pollution can influence the personal exposure to indoor air pollution, the selection of sub-cities before the selection of households is important. Thus, three representative sub-cities (Arada, Gulelle and Akaki Kality) were selected as sampling sites based on differences in altitude, socio-economic activities and population density. Arada is mainly characterized by high population density, medium traffic intensity and no industries; Gulelle has very few industries, medium traffic intensity and lower population density than Arada, whereas Akaki Kality is marked by lower population density than all the sub-cities, heavy industrial activities and high traffic congestion. Arada, Gulelle and Akaki Kality were selected as sampling sub-cities for the measurement of indoor air pollutants during the cooking of *Wot* using different fuel types in both the wet and dry seasons. Most of the time, the type of *Wot* selected for this study was prepared and consumed by low and middle income people. The households considered in this study were similar in their construction materials (wood wall, ceramic floor and roof of corrugated iron). Such houses are typical for low and middle income people in Addis Ababa and also in most Ethiopian cities. The fuel type they used and the willingness of families to allow the researcher into their house for measurement were also considered during the selection. In total, 45 homes (15 from each sub-city) were selected randomly due to a generally similar procedure in the cooking of *Wot*.

### Wot preparation techniques

Stew/sauce (*Wot*, in Amharic) is one of the main constituents of dishes for Ethiopians. From simple observation, the stews of lentil (*Misir*, in Amharic), pea (*Shiro*, in Amharic) and potato (*Dinich*, in Amharic) are most widely used stews. Powdered chilli pepper (*Berbera*, in Amharic), onion and salt are the primary ingredients used for the preparation of lentil (*Misir*), pea (*Shiro*) and potato (*Dinich*) *Wot*. Besides these, garlic, tomato paste and turmeric are optional ingredients.

The name given for each *Wot* is based on the major ingredients it contains. Thus, lentil sauce (*Misir-Wot*) contains lentil in large proportion. Similarly, *Shiro Wot* and potato sauce (*Dinich Wot*) contain a large proportion of powdered pea and chopped potato, respectively. The cuisine of such stew is similar, besides the difference in the major ingredients. Moreover, these types of stew are a popular staple food for many Ethiopians.

### PM and TVOC measurement methods

The levels of TVOC during the cooking of *Wot* using three different types of fuel were measured in 45 individual households with an easy-to-use and portable sensor (AEROQUAL series 500; Aeroqual Limited, Auckland, New Zealand) with a 2-min interval. The sampler was calibrated according to the manufacturer's procedure, at a flow rate of 0.03 m<sup>3</sup>/h (0.5 l/min). The instrument used a photoionization detector (PID). Once the sampled air is passed through the inlet, ionization of the VOCs molecules by UV absorption occurs, and the ions are counted by a detector. These ions produce a current that is amplified and displayed on the meter as 'ppm'. The instrument has a detection range of 0–1000 ppm. The measurements for both PMs and TVOC were carried out between 1 July and 30 September 2015, for the wet season and 15 November 2015 and 10 March 2016, for the dry season, in two rounds each. The two seasons (wet and dry) were selected for this study because the concentration of pollutants may depend on seasonal variations. Also, the moisture content of the biomass fuel used for baking is usually different in the wet and dry seasons. Usually, the *Wot* cook remains near to the stove until cooking is completed. Hence, the sampler was put 1 m above the ground and 1 m from the stove during cooking of *Wot*, in order to simulate the breathing zone of the cook. During the sampling time (*Wot* cooking time), all other activities were stopped to prevent their interference. The measurements were started 10 min before starting the cooking and continued until the cooking ended.

The level of PM while cooking *Wot* was measured by a portable sensor AROCET531S (Met One Instrument, Inc. Grants Pass, OR 07526, USA). The instrument contained an ambient air inlet nozzle which reduces turbulence in air sampling. It was calibrated at a flow rate of 0.1698 m<sup>3</sup>/h (2.83 l/min). It also measured the room temperature (RT) and humidity. The PM instrument used zero count test calibration, which was performed weekly. The detection limit of the instrument was 0.1 µg.m<sup>-3</sup>. The measurement of particles was based on the counts and sizes of particles in eight different size ranges, with a proprietary algorithm to convert count data to mass measurements (µg.m<sup>-3</sup>). The instrument calculated a volume for each detected particle and then assigned a standard density for the mass conversion. The standard density value was augmented by a K-Factor setting to improve measurement accuracy. The instrument provided a separate K-Factor setting for each measurement range (PM<sub>1</sub>, PM<sub>2.5</sub>, PM<sub>4</sub>, PM<sub>7</sub>, PM<sub>10</sub> and TSP). K-Factor values should be empirically derived via a comparison with a reference unit. If a reference unit is unavailable, the recommended K-Factor setting was 3.0. The instrument measured the PMs in the concentration range 0–1000 µg.m<sup>-3</sup>.

The PM sampling instrument was operated separately beside the TVOC sampling instrument and followed a similar sampling procedure.

### Statistical package for the data analysis

A series of individual concentration measurements of PM (µg.m<sup>-3</sup>) and TVOC (mg.m<sup>-3</sup>) at 2-min intervals were made in the 45 individual households. The data were classified into three categories based on the type of fuel (charcoal, kerosene and electricity). Data were analysed by statistical data analysis using IBM SPSS version 20.0, Microcal™ Origin version 6.0 (Microcal software, Inc., USA) and Microsoft Excel 2013. Initially, Shapiro-Wilks test was used for testing normality of the data. When the data were not normally distributed and concentrations were better described as log-normally distributed, the geometric mean (GOM) and geometric standard deviation (GSD) would be better for representation of data.<sup>41</sup>

The Kruskal-Wallis sample test (used to compare all data) and the Kruskal-Wallis H independent sample test (used to compare pairs) were used to evaluate the significance of differences in PM and TVOC concentrations across fuel types and within a fuel type. Finally, the Wilcoxon signed-rank test was applied to evaluate the significant differences in the concentrations of PM and TVOC between wet and dry seasons. The significance level for all tests was set to 0.05.

## Results

### Description of households

Among the selected households, 15 households used electricity, 15 used kerosene and 15 used charcoal as fuel sources for cooking stoves. The demographic data ranged from one to six individuals per household, and the kitchen volume ranged 4.36–46.9 m<sup>3</sup>. The kitchen ventilation surface areas were in the range of 0.3–2.40 m<sup>2</sup>. As far as the location of the cooking place was concerned, 12 cooking sites were in a separated area from the living room and the rest were inside the living room. Table 1 shows detailed information of the kitchen characteristics including the fuel type, ventilation type, the volume of the kitchen and family size for selected households.

### PM and TVOC levels

In this study, the concentration of PM and TVOC was measured during the cooking of *Wot* using electricity, kerosene and charcoal fuel. The Shapiro-Wilks test was applied to concentration data for each fuel type at different seasons, which showed that the data did not have a normal distribution. Hence, all concentration data are reported as GOM and GSD for both seasons. The GOM concentrations of PM and TVOC emitted while cooking using electricity, charcoal and kerosene for the wet and dry seasons are shown in Figure 1, and the data are summarized in Table 2. The room temperature (RT) and the relative humidity (RH) of the room were recorded during the cooking of *Wot*, and these results are also given in Table 2.

The GOM of PM<sub>1</sub>, PM<sub>2.5</sub>, PM<sub>4</sub>, PM<sub>7</sub>, PM<sub>10</sub>, TSP and TVOC measured for all fuels whilst cooking during the *wet* season ranged from 11.3–21.2, 25.4–45.1, 45.4–77.9, 77.0–122, 91.8–151, 109–190 and 350–1200 µg.m<sup>-3</sup>, respectively. The maximum concentrations of all types of pollutants were recorded while cooking using charcoal as a fuel source. By comparison, the minimum concentrations of PM<sub>1</sub>, PM<sub>7</sub>, PM<sub>10</sub>, TSP, TVOCs and PM<sub>2.5</sub>, PM<sub>4</sub> were recorded whilst cooking using electricity and kerosene fuels, respectively. The levels of TVOC were much higher than levels of PM emission from each fuel type.

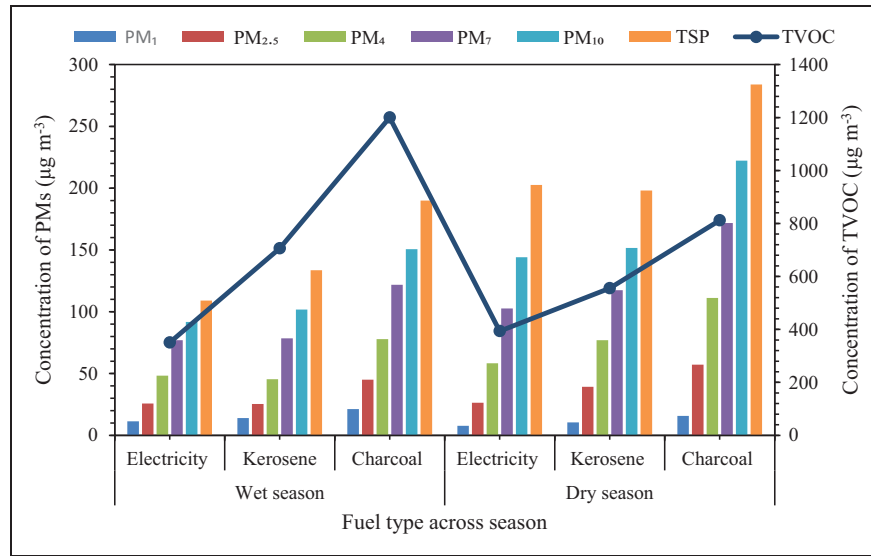
A Kruskal-Wallis test was applied to the data obtained for electricity, kerosene and charcoal fuel types, which showed that both PM and TVOC concentrations were significantly different across fuel types ( $p < 0.05$ ). However, the difference was not recognized where it occurred, i.e. is it between the electricity and kerosene or between electricity and charcoal or between kerosene and charcoal? Hence, individual pair-wise comparison was necessary to determine the source differences. Therefore, Kruskal-Wallis H



**Table 1.** Some characteristics of kitchens in the selected households at different sampling sites.

Site code	Fuel type	Ventilation type/area of ventilation (m <sup>2</sup> )/position	Kitchen position	DD	Type of sauce/ <i>Wot</i>		Kitchen volume (m <sup>3</sup> )
					Dry season	Wet season	
ar1A	Electricity	Door, 1.62, open	LR	5	<i>Shiro</i>	<i>Shiro</i>	34.8
ar1B	Electricity	Door, 1.43, open	NLR	5	<i>Shiro</i>	<i>Shiro</i>	11.1
ar1C	Electricity	Door, 1.44, open	LR	2	<i>Shiro</i>	<i>Shiro</i>	8.28
ar1D	Electricity	Door, 1.52, open	LR	5	<i>Shiro</i>	<i>Shiro</i>	20.0
ar1E	Electricity	Door, 1.62, open	NLR	4	<i>Shiro</i>	<i>Misir</i>	16.8
ar2A	Charcoal	Door, 1.2, open	LR	3	<i>Potato</i>	<i>Potato</i>	16.0
ar2B	Charcoal	Door, 1.28, open	NLR	5	<i>Shiro</i>	<i>Shiro</i>	11.5
ar2C	Charcoal	Door, 1.71, open	NLR	5	<i>Misir</i>	<i>Misir</i>	4.36
ar2D	Charcoal	Door, 1.71, open	LR	4	<i>Shiro</i>	<i>Shiro</i>	14.04
ar2E	Charcoal	Door, 1.48, open	NLR	4	<i>Shiro</i>	<i>Shiro</i>	9.01
ar3A	Kerosene	Door, 1.4, open	NLR	2	<i>Shiro</i>	<i>Shiro</i>	8.28
ar3B	Kerosene	Door, 1.57, open	LR	1	<i>Shiro</i>	<i>Shiro</i>	19.0
ar3C	Kerosene	Door, 1.67, open	LR	1	<i>Shiro</i>	<i>Shiro</i>	21.5
ar3D	Kerosene	Door, 1.67, open	LR	1	<i>Misir</i>	<i>Misir</i>	18.8
ar3E	Kerosene	Door, 1.2, open	NLR	3	<i>Potato</i>	<i>Potato</i>	14.5
gu1A	Electricity	Door, 1.6, partially open	LR	5	<i>Shiro</i>	<i>Shiro</i>	18.7
gu1B	Electricity	Door, 1.81, closed	LR	2	<i>Shiro</i>	<i>Shiro</i>	17.5
gu1C	Electricity	Door, 1.85, open	NLR	6	<i>Shiro</i>	<i>Shiro</i>	16.6
gu1D	Electricity	Door, 1.66, open	LR	1	<i>Shiro</i>	<i>Shiro</i>	8.34
gu1E	Electricity	Door, 1.2, open	LR	3	<i>Shiro</i>	<i>Misir</i>	16.0
gu2A	Charcoal	Door, 1.81, open	LR	6	<i>Shiro</i>	<i>Misir</i>	13.7
gu2B	Charcoal	Door, 1.86, open	LR	4	<i>Shiro</i>	<i>Potato</i>	22.9
gu2C	Charcoal	Door, 1.00, open	NLR	4	<i>Shiro</i>	<i>Shiro</i>	15.8
gu2D	Charcoal	Door, 1.95, open	NLR	4	<i>Misir</i>	<i>Shiro</i>	12.0
gu2E	Charcoal	Door, 1.57, open	LR	3	<i>Misir</i>	<i>Shiro</i>	16.4
gu3A	Kerosene	Door, 1.71, open	LR	3	<i>Misir</i>	<i>Shiro</i>	17.4
gu3B	Kerosene	Door, 1.76, open	LR	4	<i>Shiro</i>	<i>Misir</i>	21.0
gu3C	Kerosene	Door, 1.31, partially open	LR	5	<i>Potato</i>	<i>Shiro</i>	14.8
gu3D	Kerosene	Door, 1.71, open	LR	2	<i>Shiro</i>	<i>Shiro</i>	9.61
gu3E	Kerosene	Door, 1.57, partially open and window, 0.26, open	LR	5	<i>Shiro</i>	<i>Shiro</i>	12.4
ak1A	Electricity	Door, 1.2, open	LR	3	<i>Shiro</i>	<i>Shiro</i>	16.0
ak1B	Electricity	Door, 1.94, open	LR	2	<i>Shiro</i>	<i>Shiro</i>	9.26
ak1C	Electricity	Door, 1.76, open	LR	5	<i>Shiro</i>	<i>Misir</i>	44.8
ak1D	Electricity	Door, 1.95, open	LR	4	<i>Potato</i>	<i>Potato</i>	12.5
ak1E	Electricity	Door, 1.67, partially open and window, 0.3, open	LR	2	<i>Shiro</i>	<i>Misir</i>	5.32
ak2A	Charcoal	Door, 1.76, open	LR	2	<i>Shiro</i>	<i>Shiro</i>	18.7
ak2B	Charcoal	Door, 1.11, open	LR	6	<i>Shiro</i>	<i>Shiro</i>	5.00
ak2C	Charcoal	Door, 1.85, open	LR	3	<i>Shiro</i>	<i>Shiro</i>	21.0
ak2D	Charcoal	Door, 1.95, open	LR	5	<i>Misir</i>	<i>Shiro</i>	32.1
ak2E	Charcoal	Door, 2.4, open	NLR	4	<i>Misir</i>	<i>Misir</i>	42.0
ak3A	Kerosene	Door, 1.76, open	LR	3	<i>Shiro</i>	<i>Shiro</i>	7.98
ak3B	Kerosene	Door, 1.67, open	LR	4	<i>Misir</i>	<i>Misir</i>	15.6
ak3C	Kerosene	Door, 1.71, open	LR	6	<i>Misir</i>	<i>Shiro</i>	35.9
ak3D	Kerosene	Door, 1.71, open	LR	4	<i>Shiro</i>	<i>Shiro</i>	46.9
ak3E	Kerosene	Door, 1.62, open	NLR	2	<i>Shiro</i>	<i>Shiro</i>	4.00

Site code: ar: Arada sub-city; gu: Gullele sub-city; ak: Akaki Kality sub-city; 1, 2 and 3: indicates fuel types electricity, charcoal and kerosene, respectively; A, B, C, D and E represent five different households in a particular each sub-city; LR: kitchen found in living room; NLR: kitchen found in separate from living room; DD: demographic data and the number indicates number of members of the family; K-volume: kitchen volume. The wet and dry seasons indicate the time of measurements; wet season: between 1 July and 30 September 2015 and dry season: between 15 November 2015 and 10 March 2016.



**Figure 1.** The levels of PM and TVOC during the cooking of *Wot* using different fuel types in dry and wet seasons. PM: particulate matter; TVOC: total volatile organic compound.

**Table 2.** The GOM concentration of PM and TVOC and mean values of RT (room temperature), RH (relative humidity of the room) and time period during the cooking of *Wot* using different types of fuels during the wet and dry seasons.

Season type	Fuel type	PM <sub>1</sub> (µg.m <sup>-3</sup> )	PM <sub>2.5</sub> (µg.m <sup>-3</sup> )	PM <sub>4</sub> (µg.m <sup>-3</sup> )	PM <sub>7</sub> (µg.m <sup>-3</sup> )	PM <sub>10</sub> (µg.m <sup>-3</sup> )	TSP (µg.m <sup>-3</sup> )	TVOC (µg.m <sup>-3</sup> )	RT (°C)	RH (%)	Time (min)
Wet	Electricity	11.4 ± 1.96	25.8 ± 2.18	48.4 ± 2.23	77.0 ± 2.21	91.8 ± 2.21	109 ± 2.17	350 ± 2.08	21.3 ± 1.13	60.6 ± 1.14	78.0 ± 1.60
	Kerosene	14.0 ± 1.74	25.4 ± 1.70	45.4 ± 1.68	78.5 ± 1.65	102 ± 1.64	134 ± 1.66	706 ± 2.66	22.5 ± 1.10	57.7 ± 1.20	96.0 ± 1.42
	Charcoal	21.2 ± 1.85	45.1 ± 2.25	77.9 ± 2.35	122 ± 2.23	151 ± 2.17	190 ± 2.13	1200 ± 2.17	21.5 ± 1.11	57.4 ± 1.16	103 ± 1.23
Dry	Electricity	7.68 ± 1.63	26.4 ± 2.02	58.3 ± 1.98	103 ± 1.89	144 ± 1.87	203 ± 1.84	394 ± 1.96	22.3 ± 1.16	52.9 ± 1.25	89.4 ± 1.40
	Kerosene	10.5 ± 1.88	39.2 ± 2.48	76.9 ± 2.59	117 ± 2.47	152 ± 2.37	198 ± 2.28	555 ± 2.14	21.0 ± 1.12	52.9 ± 2.29	86.2 ± 1.28
	Charcoal	15.6 ± 1.66	57.2 ± 1.89	111 ± 1.79	172 ± 1.68	222 ± 1.62	284 ± 1.58	812 ± 1.96	21.1 ± 1.11	50.1 ± 1.20	91.4 ± 1.35

independent sample test was further applied by case selection. Thus, charcoal fuel was compared to electricity and kerosene fuels separately, and the result showed a significant difference in the concentration of emitted PM and TVOC ( $p < 0.05$ ). Likewise, electricity and kerosene fuels were compared for the PM and TVOC concentrations, which showed a significant difference in their concentration ( $p < 0.05$ ), but PM<sub>1</sub>, PM<sub>7</sub> and PM<sub>10</sub> did not show a significant difference ( $p > 0.12$ ). Moreover, the average time taken for the cooking of *Wot* at the wet season using electricity, kerosene and charcoal fuels were 78.0, 96.0 and 103 min, respectively. Difference in the cooking time of *Wot* using electricity, kerosene and charcoal fuels might have also contributed to the differences in the levels of PMs and TVOC.

The GOM concentrations of PM<sub>1</sub>, PM<sub>2.5</sub>, PM<sub>4</sub>, PM<sub>7</sub>, PM<sub>10</sub>, TSP and TVOC for all fuels during the *dry* season were 7.68–15.9, 26.4–57.2, 58.3–111, 103–172, 144–222, 198–284 and 294–812 µg.m<sup>-3</sup>, respectively. For both PM and TVOC pollutants, the lowest concentrations were found using electricity fuel

(except for TSP while using kerosene fuel). The highest concentration of TSP was found using charcoal fuel. As observed in the wet season, levels of TVOC in the dry season were also much higher than levels of emission of PMs from each fuel type. This showed that the general trend of emission of PMs and TVOC is the same in both the wet and dry seasons. The average cooking times of *Wot* using electricity, kerosene and charcoal fuel during the dry season were 89.4, 86.2 and 91.4 min, respectively. This shows that average cooking times of *Wot* were slightly shorter in the dry season as compared to the wet season.

Similar to wet season, a Kruskal-Wallis test was applied to the data for electricity, charcoal and kerosene fuel in the dry season, and this showed a significant difference in GOM concentrations of PM and TVOC across these three fuel types ( $p < 0.05$ ). A Kruskal-Wallis H independent sample test was also carried out, such that charcoal was compared to kerosene and electricity fuels separately. This showed a significant difference in the PM and TVOC concentration

( $p < 0.05$ ). Similarly, data for electricity and kerosene were compared, which showed a significant difference in both PM and TVOC concentrations with  $p < 0.05$ , except PM<sub>10</sub> and TSP which did not show a significant difference ( $p > 0.199$ ).

In general, as shown in Figure 1, the trend for GOM concentrations of PM during the cooking of *Wot* was in the order of charcoal > kerosene > electricity fuel during the dry season, whereas the emission pattern shifted to charcoal > kerosene  $\approx$  electricity fuel, for the wet season. The patterns of TVOC were in the order of charcoal > kerosene > electricity fuel for both seasons. The variation in the concentration of PM and TVOC might be due to differences in the temperature and humidity of the room. Also, the moisture content of the charcoal fuel, the type of wood species that the charcoal was made from and the condition of the fire during measurements might be contributing to the concentration difference in PMs and TVOC in the wet and dry seasons using charcoal fuel. Also, the variation could be due to different types of ventilation, the position of the stoves from the ventilation location, the volume of the kitchen and the dispersion and deposition rates of different particles.<sup>42–44</sup>

### Temporal variation of PM and TVOC

The temporal variations in the GOM concentrations of PM and TVOC were calculated using a Wilcoxon signed-rank test for electricity, kerosene and charcoal fuels, and the results are given in Table 3. The Wilcoxon signed-rank test showed a significant difference in the concentration of PM and TVOC for nearly all fuel types between the dry and wet season at  $p < 0.05$ . However, PM<sub>2.5</sub> and TVOC did not show a

significant difference for electricity as fuel. Although there is no difference in the cooking RT, the relative humidity of the cooking room and the cooking time of *Wot* showed a significant difference with seasons, which might affect resuspension rate or the dispersion of pollutants for the dry and wet season with fuel types. Consequently, this variation might contribute to the changes in concentrations of PM and TVOC between seasons. Also, the moisture content of the fuel, status of the fire and the formation of secondary pollutants (through condensation and coagulation) during measurements might also result in differences in PM and TVOC concentration. Another reason for the difference in the concentration of PM in wet and dry seasons could be due to the formation of secondary pollutants, which can be formed with pollutants released from the cooking process itself, without the involvement of any other pollutants emitted from other sources.<sup>45,46</sup>

Figure 2 shows the time-patterns of PM and TVOC concentrations during the cooking of *Wot* using electricity (ar1A), charcoal (ar2A) and kerosene (ar3A) fuels at Arada sub-city. The concentration of each pollutant was constant before cooking (for about 10 min) because the stoves were off and became elevated once the stoves were turned on for the cooking. This might be due to the fact that the level of pollutants during cooking varied with the status of fire (whether it was off, starting, burning or smouldering) and adding or removing fuel, the type of food and fuel. Previous studies have shown that the concentrations of pollutants become high when starting cooking using biomass fuel, afterwards with smouldering of the fuel to make a better flame, adding more biomass fuel and also due to frying of the food.<sup>43,47–49</sup>

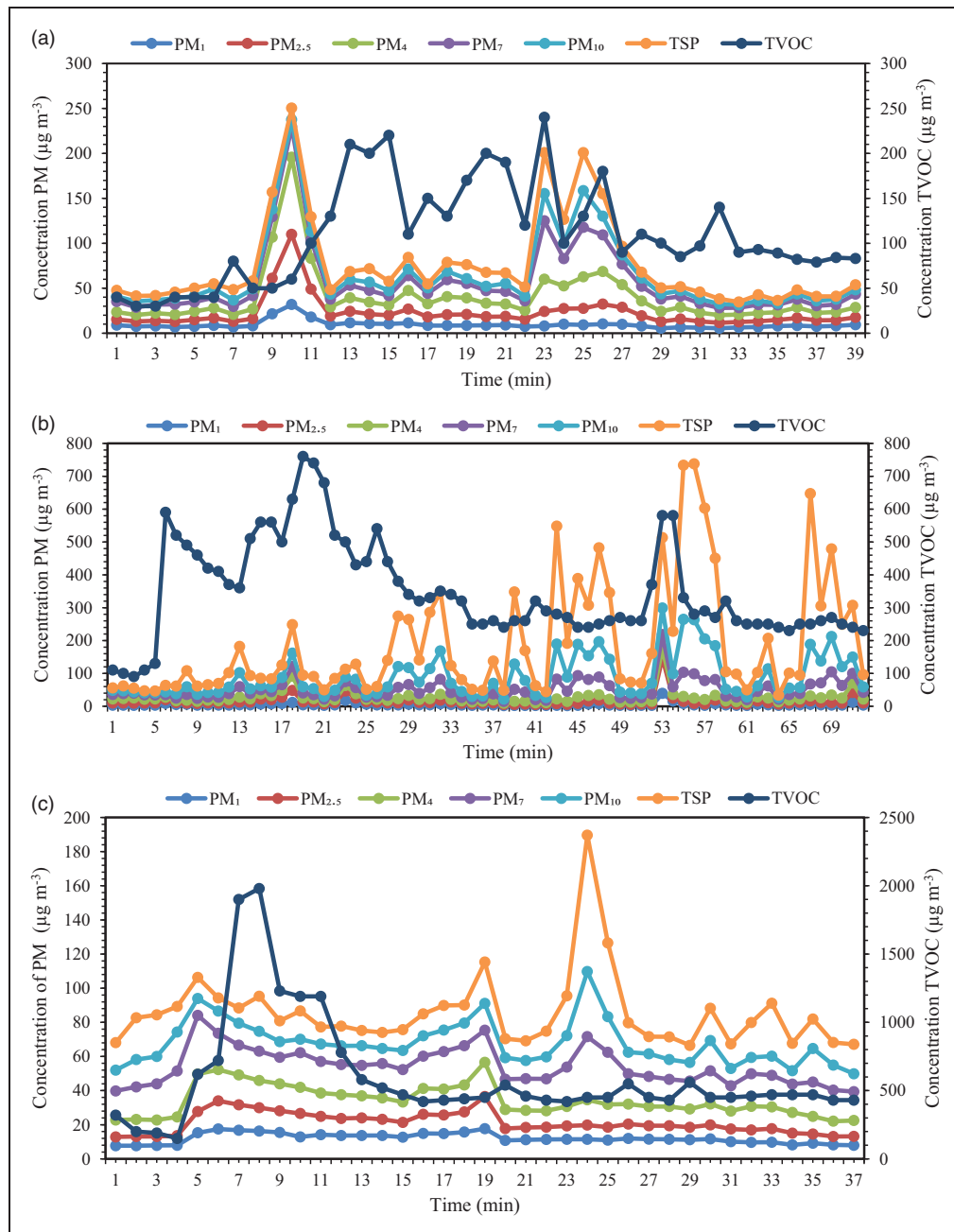
**Table 3.** Wilcoxon signed rank test result between the dry and wet season for emissions of PM, TVOC, RT and RH whilst cooking using electricity, kerosene and charcoal fuels.

	Electricity		Charcoal		Kerosene	
	Z	Asymp. sig. (two-tailed)	Z	Asymp. sig. (two-tailed)	Z	Asymp. sig. (two-tailed)
PM <sub>1</sub> during dry–PM <sub>1</sub> during wet season	-8.01 <sup>a</sup>	0.000	-12.5 <sup>a</sup>	0.000	-7.02 <sup>a</sup>	0.000
PM <sub>2.5</sub> during dry–PM <sub>2.5</sub> during wet season	-0.004 <sup>b</sup>	0.997	-4.53 <sup>b</sup>	0.000	-11.8 <sup>b</sup>	0.000
PM <sub>4</sub> during dry–PM <sub>4</sub> during wet season	-2.70 <sup>b</sup>	0.007	-7.78 <sup>b</sup>	0.000	-13.3 <sup>b</sup>	0.000
PM <sub>7</sub> during dry–PM <sub>7</sub> during wet season	-4.54 <sup>b</sup>	0.000	-7.96 <sup>b</sup>	0.000	-11.2 <sup>b</sup>	0.000
PM <sub>10</sub> during dry–PM <sub>10</sub> during wet season	-8.43 <sup>b</sup>	0.000	-8.98 <sup>b</sup>	0.000	-11.5 <sup>b</sup>	0.000
TSP during dry–TSP during wet season	-12.3 <sup>b</sup>	0.000	-9.27 <sup>b</sup>	0.000	-11.6 <sup>b</sup>	0.000
TVOC during dry–TVOC during wet season	-0.99 <sup>a</sup>	0.321	-7.93 <sup>a</sup>	0.000	-5.86 <sup>a</sup>	0.000
RT during dry–RT during wet season	-6.99 <sup>b</sup>	0.000	-2.33 <sup>a</sup>	0.020	-13.1 <sup>a</sup>	0.000
RH during–RH during wet season	-14.7 <sup>a</sup>	0.000	-12.3 <sup>a</sup>	0.000	-9.28 <sup>b</sup>	0.000

RH: relative humidity; RT: room temperature; TVOC: total volatile organic compound; TSP: total suspended particulate matter; PM: particulate matter.

<sup>a</sup>Positive ranks.

<sup>b</sup>Negative ranks.



**Figure 2.** A typical time-pattern concentration of (i) PM and (ii) TVOC during cooking of *Wot* in Arada sub-city using electricity (a), charcoal (b) and kerosene (c) fuels during the wet season. PM: particulate matter; TVOC: total volatile organic compound; TSP: total suspended particulate matter.

**Discussion**

In this study, we measured the indoor level of pollutants before and during cooking. We did not measure outdoor levels since the focus was to measure the pollutants in the indoor air during cooking.

Exposure assessment for different microenvironments is needed for estimating the health impact of the exposed person and to determine appropriate

remedial interventions. Hence, this study is the first attempt to look at the indoor air pollution during the cooking of *Wot* using energy sources of electricity fuel, kerosene fuel and charcoal fuel at different microenvironments. The general trend showed the level of PM and TVOC found in the order of electricity < kerosene < charcoal fuel. The overall GOM of PM and TVOC regardless of the season is given in Table 4. The highest GOM concentration of PM and TVOC



**Table 4.** The overall GOM concentration of PMs, TVOC, time, K-volume, RT and RH measured during cooking using different types of fuels.

Fuel type	PM <sub>1</sub> ( $\mu\text{g.m}^{-3}$ )	PM <sub>2.5</sub> ( $\mu\text{g.m}^{-3}$ )	PM <sub>4</sub> ( $\mu\text{g.m}^{-3}$ )	PM <sub>7</sub> ( $\mu\text{g.m}^{-3}$ )	PM <sub>10</sub> ( $\mu\text{g.m}^{-3}$ )	TSP ( $\mu\text{g.m}^{-3}$ )	TVOC ( $\mu\text{g.m}^{-3}$ )	RT (°C)	RH (%)	Time (min)	Kitchen volume (m <sup>3</sup> )
Electricity	9.20±1.86	26.1±2.20	53.5±2.01	90.0±1.96	117±1.97	152±2.00	373±2.02	21.9±1.13	56.3±1.48	84.0±1.48	14.7±1.59
Charcoal	18.6±1.78	50.2±2.16	91.4±2.23	142±2.12	179±2.09	228±2.06	1006±2.32	21.3±1.13	54.0±1.22	97.7±1.32	14.9±1.74
Kerosene	12.2±1.84	31.4±2.16	58.7±2.24	95.5±2.12	124±2.07	162±2.03	628±2.43	21.7±1.12	50.2±1.25	91.1±1.36	16.2±1.76

RH: relative humidity; RT: room temperature; TVOC: total volatile organic compound; TSP: total suspended particulate matter; PM: particulate matter.

**Table 5.** Comparison of the percentage emission exposure reduction to PM and TVOC by using electricity and kerosene fuel for cooking instead of using charcoal fuel; and electricity fuel instead of kerosene fuel for cooking.

PM <sub>1</sub> (%)	PM <sub>2.5</sub> (%)	PM <sub>4</sub> (%)	PM <sub>7</sub> (%)	PM <sub>10</sub> (%)	TSP (%)	TVOC (%)	Remarks
50.4	48.0	41.5	36.8	34.7	33.1	62.9	Electricity vs. charcoal
34.4	37.4	35.8	32.9	31.1	28.9	37.6	Kerosene vs. charcoal
24.4	16.9	8.79	5.80	5.23	5.82	40.6	Electricity vs. kerosene

TVOC: total volatile organic compound; TSP: total suspended particulate matter; PM: particulate matter.

was recorded whilst cooking using charcoal, whereas the lowest values were measured using electricity.

Even though recent studies have shown exposure to high levels of PM for a short time in the order of one or more hour can cause serious health risks, guidelines appropriate to stove emissions are not available.<sup>36,37</sup>

Furthermore, studies have shown that the use of clean fuels, such as electricity and kerosene instead of biomass fuel could reduce the total exposure while preparing foods. The amount of pollutant reduction depends on the type of food prepared, the type of ventilation (area, position either fully open or partially open), the location of the stove from the ventilation position in the kitchen and the type of biomass fuel used.<sup>42,43,50–52</sup> This study presents the concentration variations measured during the cooking of an Ethiopian traditional *Wot* using the most common fuels namely electricity, kerosene and charcoal. Table 5 shows the percentage reduction of emission exposure using electricity and kerosene fuels as compared to charcoal fuel during the preparation of *Wot*. Similarly, emission exposure reduction using electricity fuel as compared to kerosene fuel was also estimated. Equations (1) and (2) were used for these calculations.

$$\begin{aligned} \text{\%Emission exposure reduction} \\ = \frac{(\text{PCF} - \text{PEF or PKF})}{\text{PCF}} \times 100 \end{aligned} \quad (1)$$

$$\text{\%Emission exposure reduction} = \frac{(\text{PKF} - \text{PEF})}{\text{PKF}} \times 100 \quad (2)$$

Where PCF is the overall geometric mean concentration of pollutant released during cooking using charcoal fuel measured during both dry and wet seasons; PKF is the total geometric mean concentration of pollutants released during cooking using kerosene fuel measured during both dry and wet seasons; PEF is the overall geometric mean concentration of pollutants released during cooking using electricity fuel measured during both dry and wet seasons.

Overall, Table 5 shows that cooking *Wot* using electricity instead of charcoal can reduce emission exposure by 33–63%. Likewise, a *Wot* cooker can reduce emission exposure by 29–38% by using kerosene instead of charcoal. Lastly, using electricity instead of kerosene can reduce exposure to PM and TVOC pollutants by 5–41%.

Table 6 compares the data of the present study to research in other countries on levels of indoor air pollutants in kitchens or during cooking using different types of fuels. The particulate concentrations recorded in the present study are comparable with some studies<sup>11,43,50–52</sup> using kerosene (India, Andhra Pradesh and Bangladesh), whereas most studies<sup>43,50–53</sup> reported much higher PM than the present study, using biomass fuel such as charcoal, wood and dung for cooking.

Furthermore, in this study, the health risk assessment due to the exposure to PM<sub>2.5</sub>, PM<sub>10</sub> and TSP was considered. Human health risk assessment due to air contaminants depends on the type of pollutants and the extent of exposure. Hazard identification, dose–response assessment, exposure assessment and risk characterization are the main steps used in a health risk assessment.<sup>54–56</sup> Each of these steps was used to estimate the health risk due to PM emitted during the

**Table 6.** Summary of indoor air pollution in developing countries using different fuel types.

Country	Description	Particulate concentration range ( $\mu\text{g}\cdot\text{m}^{-3}$ )
Bangladesh <sup>52</sup>	Aerial 24 h	PM <sub>10</sub> : use of dung (291); firewood (263); sawdust (237); natural gas stove (101); kerosene stove (134)
Southern India <sup>43</sup>	Aerial 24 h	RSP: wood (204); wood chips (266); agricultural produce (246); kerosene (78); gas (78)
Nigeria, Ibadan <sup>11</sup>	Twice a day (morning: 7–9 am and evening: 5–7 pm)	PM <sub>10</sub> : use of kerosene (231–365)
India, Andhra Pradesh <sup>51</sup>	24 h kitchen concentration	RPM (GOM): dung (470); wood (340); kerosene (156) and gas (61)
Dominican Republic <sup>53</sup>	Half in kitchen and half in the child's bedroom for 10 h	RPM (GOM): charcoal (27.9 l); gas (17.6)
India, Delhi <sup>50</sup>	Kitchen during cooking	RPM: LPG (890), kerosene (690), wood (1370) and coal (1090)
Ethiopia, Addis Ababa (This study)	During cooking <i>Wot</i> during wet season	PM <sub>1</sub> , PM <sub>2.5</sub> , PM <sub>4</sub> , PM <sub>7</sub> , PM <sub>10</sub> , TSP and TVOC (GOM): electricity (11.3, 25.7, 48.4, 77.0, 91.8, 109 and 350); charcoal (21.2, 45.1, 77.9, 122, 151, 190 and 1200); kerosene (14.0, 25.4, 45.4, 78.5, 102, 134 and 706)
	During cooking <i>Wot</i> during dry season	electricity (7.68, 26.4, 58.3, 103, 144, 203 and 394); charcoal (15.6, 57.2, 111, 172, 222, 283 and 812); kerosene (10.5, 39.2, 76.9, 117, 152, 198 and 555)

Note: RPM is respirable particulate matter.

cooking of *Wot*. Although all PM were identified as hazardous, only PM<sub>2.5</sub>, PM<sub>10</sub> and TSP were considered due to availability of toxicity data. These data include the reference dose (RfD) values for pollutants which were used in the calculation of a hazard quotient (HQ), where HQ is the ratio of the potential exposure to a substance and the level at which no adverse effects are expected. In the second step, average daily intake (called chronic daily intake ADD ( $\mu\text{g}\cdot\text{kg}^{-1}\cdot\text{day}^{-1}$ )) was calculated using equations (3) and (4). The annual threshold values set by the World Health Organization (WHO)<sup>27</sup> and the US EPA<sup>57</sup> were used as the RfD in the dose–response step. Finally, the HQ was calculated and used in the risk characterization for different microenvironments during the cooking of *Wot* using charcoal, kerosene and electricity.<sup>54–56</sup>

In Ethiopia, women take responsibility for cooking the *Wot* and feeding the family. They usually start cooking at the age of 17 and continue until they are retired. Hence, only women's health impacts were estimated in this study. The current life expectancy for an Ethiopian woman is around 67, and so the total exposure duration of 50 years was taken for the calculation of HQ and ADD. The exposure frequency of 365 days/year was also assumed. The air intake rate for an adult is 20 m<sup>3</sup> per day, and the body weight for adult Ethiopian women was assumed as the African adult weight of 60.7 kg.<sup>57,58</sup> The times taken for the cooking of *Wot* were recorded during measurement of PM, and these were then doubled to estimate exposure times since the cooking was done twice a day. Thus, 2.8, 3.26 and 3.04 h.day<sup>-1</sup> were

used as exposure times for electricity, charcoal and kerosene fuel, respectively. The details considered in the HQ calculation are summarized in Table 7.

The exposure duration of this study is 2129, 2479 and 2312 days for the measurement of pollutants concentrations during cooking using electricity, charcoal and kerosene fuels, respectively, which are equivalents to 5.83, 6.79 and 6.33 years, respectively. The US EPA<sup>57</sup> classifies an exposure as acute for an exposure duration below two weeks, as sub-chronic exposure when the duration is above two weeks and below seven years, and as chronic for exposure duration above seven years. Thus, the exposure duration of this study falls in the category 'above two weeks and below seven years'. Hence, it is classified under the sub-chronic type of exposure. Hence, RfDs for sub-chronic exposure are used to calculate HQ, which are the annual averages of PM<sub>2.5</sub> (10  $\mu\text{g}\cdot\text{m}^{-3}$ ), PM<sub>10</sub> (20  $\mu\text{g}\cdot\text{m}^{-3}$ ) and TSP (150  $\mu\text{g}\cdot\text{m}^{-3}$ ) values set by WHO and US EPA.<sup>26,27</sup> The overall geometric means of PM<sub>2.5</sub>, PM<sub>10</sub> and TSP (regardless of seasons) have been used for health risk assessments.

Equations (3) and (4) were used for the calculation of ADD and HQ to characterize the health risk due to PM<sub>2.5</sub> and PM<sub>10</sub>.<sup>54,55</sup>

$$ADD = \frac{CA \times IR \times ED}{AT \times BW} \quad (3)$$

$$ED = \frac{EF \times ET \times DE \times 1 \text{ day}}{24 \text{ h}} \quad (4)$$

**Table 7.** Exposure time, exposure frequency, duration exposure, exposure duration, averaging time (AT), intake rate (IR) and body weight (BW) for adult woman.

Fuel type	Exposure time (h.day <sup>-1</sup> )	Exposure frequency (days.year <sup>-1</sup> )	Duration exposure (years)	Exposure duration (days)	AT (days)	IR (m <sup>3</sup> .day.y <sup>-1</sup> )	BW (kg)
Electricity	2.8	365	50	2129	18,250	20	60.7
Charcoal	3.26	365	50	2479	18,250	20	60.7
Kerosene	3.04	365	50	2312	18,250	20	60.7

where *ADD* is the average daily intake ( $\mu\text{g}\cdot\text{kg}^{-1}\cdot\text{day}^{-1}$ ), *CA* is the contaminant concentration in air ( $\mu\text{g}\cdot\text{m}^{-3}$ ), *ET* is the exposure time spent (hours.day<sup>-1</sup>), *EF* is the exposure frequency (days.year<sup>-1</sup>), *IR* is the intake rate ( $\text{m}^3\cdot\text{day}^{-1}$ ), *ED* is the exposure duration (days), *DE* is the duration of exposure (years), *AT* is the averaging time and (*AT* (days) = *DE* (years) × 365 days.year<sup>-1</sup>) and *HQ* is defined by equation (5).

$$HQ = \frac{ADD}{RfD} \quad (5)$$

Where *HQ* is the hazard quotient; *EC* is the exposure concentration obtained from equation (1); *RfD* is the reference dose (converted to the same unit as *ADD* ( $\mu\text{g}\cdot\text{kg}^{-1}\cdot\text{day}^{-1}$ ), that *RfD* expressed in  $\mu\text{g}\cdot\text{kg}^{-1}\cdot\text{day}^{-1}$  would be equal to the *RfD* in  $\mu\text{g}\cdot\text{m}^{-3}$  multiplied by 20 m<sup>3</sup> air inhaled per person per day divided by 70 kg per person). The average weight of an African woman was assumed to be 60.7 kg, but since there is no conversion factor using this weight for calculating US EPA's *RfD* values, the weight of the person was taken as 70 kg for this study.

According to the US EPA,<sup>57</sup> if the calculated *HQ* is less than one, no adverse health effects are expected as a result of the exposure. If the calculated *HQ* value is more than one, long-term exposure to air pollution may cause significant human health effects, including loss of lung capacity and accelerated ageing of the lungs. It is also possible that individuals could develop diseases such as asthma, bronchitis, emphysema and cancer.<sup>27</sup> Hence, as seen in Table 8, using charcoal fuel alone can have a likelihood to induce health impacts for the exposed person from exposure to PM<sub>10</sub>, since its *HQ* value was larger than one. By comparison, the *HQ* values from the use of kerosene and electricity were below one for both PM<sub>2.5</sub> and PM<sub>10</sub>, which indicates that the risks of health problems using these stoves are lower than for charcoal. However, this does not mean that their contribution to the total chronic exposure is low. Thus, cooking *Wot* using electricity or kerosene fuels can contribute to the total chronic intake of PM<sub>2.5</sub> and PM<sub>10</sub> about 30–101% and 11–17% for TSP. The highest and the lowest per

cent contribution to the total chronic intake for all pollutants were due to the use of charcoal and electricity fuel, respectively.

In addition to using *HQ* for health risk assessment, the *HI*, which is the sum of *HQs* of each pollutant, has been used to assess the cumulative effect of exposure to pollutants.<sup>27</sup> The US EPA<sup>57</sup> considers that if *HI* is greater than one, the exposed person can have health problems, whereas if *HI* is less than one, they may not face serious health problems.<sup>27</sup> Thus, the *HI* was calculated to determine the risk in the use of electricity, kerosene and charcoal as fuels. The *HQ* of PM<sub>2.5</sub>, PM<sub>10</sub> and TSP was 1.06, 1.16 and 1.74, respectively. Thus, although individual pollutants (except PM<sub>10</sub> during charcoal use) may not impose health impact to a woman who is cooking, their cumulative effect has a likelihood to induce non-carcinogenic health problems.

Cooking *Wot* is the most wide and daily practice in an Ethiopian home. However, preparing food using traditional stoves can increase the individual long-term exposure to PM and TVOC. Based on this study, a person who uses charcoal fuel for extended periods could face various health-related problems due to long-term exposure to PM.

In addition, urban firewood and charcoal demand have been destructive to forestry and caused environmental damage. Ethiopia generates more than 80% of its electricity from sustainable sources such as hydro-electricity. The government should make affordable electric stoves readily available to the community. However, electrical services are irregular and often do not reach the poor urban sectors. Further research should consider the adverse health, economic and environmental effects related to cooking along with the measurement of other pollutants. Little research has been done on the possible interactions of indoor air pollution due to cooking smoke and related health outcomes and pathogenesis of respiratory diseases among the lower socioeconomic sectors.

Although this study has provided information related to the exposure level due to the PM and TVOC during the cooking of *Wot*, it has some shortcomings. First, measuring the pollutant levels together with

**Table 8.** ADD for PMs and TVOC and HQ for PM<sub>2.5</sub> and PM<sub>10</sub>.

Average daily intake across fuel type	PM <sub>1</sub>	PM <sub>2.5</sub>	PM <sub>4</sub>	PM <sub>7</sub>	PM <sub>10</sub>	TSP	TVOC
ADD for PMs and TVOC ( $\mu\text{g.kg}^{-1}.\text{day}^{-1}$ )							
ADD <sub>ele</sub>	0.34	0.96	1.97	3.32	4.32	5.62	13.8
ADD <sub>ch</sub>	0.68	1.85	3.37	5.24	6.61	8.39	37.1
ADD <sub>ker</sub>	0.45	1.16	2.16	3.52	4.56	5.96	23.2
RfD annual	–	3.24	–	–	6.56	49.4	–
HQ for PM <sub>2.5</sub> PM <sub>10</sub> and TSP							
HQ <sub>ele</sub>	–	0.29	–	–	0.66	0.11	–
HQ <sub>ch</sub>	–	0.56	–	–	1.01	0.17	–
HQ <sub>ker</sub>	–	0.35	–	–	0.69	0.12	–

ADD<sub>ele</sub>: average daily intake in using electricity fuel; ADD<sub>ch</sub>: average daily intake using charcoal fuel; ADD<sub>ker</sub>: average daily intake using kerosene fuel; HQ<sub>ele</sub>: hazard quotient in using electricity fuel; HQ<sub>ch</sub>: hazard quotient in using charcoal fuel; HQ<sub>ker</sub>: hazard quotient in using kerosene fuel; TVOC: total volatile organic compound; TSP: total suspended particulate matter; PM: particulate matter.

examination of different health outcomes was not performed due to the lack of a professional health worker (a medical technician or nurse). However, the results of this study could be used as a benchmark for future research on the examination of different health issues caused by indoor air pollution. Second, the amount and the type of chemical substances present in the smoke of charcoal fuel depend on the moisture content of the charcoal, the species of plant from which the charcoal originated and amount of charcoal used.<sup>59–61</sup> Also, the chemical constituent of the smoke depends on the status of fire (such as smouldering compared to a good flaming condition). However, this study did not consider these aspects. Third, biomass fuel smoke can emit different pollutants other than PM and TVOC. TVOC measurement does not distinguish the individual levels of VOCs, such as low-volatility organics including polynuclear aromatic hydrocarbons. Further research should be carried out on the health aspects of chronic exposure to Ethiopian traditional cooking practices along with the measurement of different pollutants levels.

However, despite these limitations, this study has demonstrated the high level of exposure to selected pollutants when cooking *Wot* using charcoal.

## Conclusion

This study is the first attempt to look at the indoor air pollution during the cooking of *Wot* using energy sources of electricity, kerosene and charcoal in different microenvironments. The general trend showed the level of PM and TVOC found in the order of electricity < kerosene < charcoal fuel. The highest GOM concentration of PM and TVOC was recorded using charcoal, whereas the lowest values were measured

using electricity. Among the three fuel types, electricity was much better in the reduction of emissions to PM and TVOC. The health risk assessment indicated that using charcoal fuel alone had a likelihood to induce health impacts of women due to exposure to PM<sub>10</sub> during cooking, since its HQ value was greater than one. By comparison, the HQ values for both PM<sub>2.5</sub> and PM<sub>10</sub> were below one when kerosene and electricity are used, which implies the risks of health problems using these stoves is lower. However, cooking *Wot* using electricity or kerosene fuels does not mean that their contribution to the total chronic exposure is low. The health risk assessment based on the HI values showed that any of the three fuel types could contribute to non-carcinogenic health problems to the *Wot* cooker due to the cumulative effect of PM<sub>2.5</sub>, PM<sub>10</sub> and TSP, although individual pollutants (except PM<sub>10</sub> during charcoal use) may not impose health impact to a cooker. Using charcoal fuel alone, the person cooking appears likely to experience adverse health impacts due to PM<sub>10</sub>, since its HQ value was larger than one.

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## Authors' contribution

The first author collected the data, conducted data analysis and primarily drafted this paper. The second author designed the research with all the required technical elements. The third author also helped in designing the research, contributed to the data analysis and critically reviewed and edited the article. The fourth author provided the devices, contributed to data analysis and revision of the article.




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### References

- Anenberg SC, Balakrishnan K, Jetter J, Masera O, Mehta S, Moss J and Ramanathan V. Cleaner cooking solutions to achieve health, climate, and economic co-benefits. *Environ Sci Technol* 2013; 47: 3944–3952.
- Albalak R, Bruce N, McCracken JP, Smith KR and Gallardo TD. Indoor respirable particulate matter concentrations from an open fire, improved cookstove, and LPG/open fire combination in a rural guatemalan community. *Environ Sci Technol* 2001; 35: 2650–2655.
- Biruck D, Suleiman H and Araya A. Household fuel use and acute respiratory infections among younger children: an exposure assessment in Shebedino Wereda, Southern Ethiopia. *Afr J Health Sci* 2011; 18: 31–36.
- Hassena AA, Kebedeb SB and Wihib NM. Design and manufacturing of thermal energy based Injera baking glass pan. *Energy Proc* 2016; 93: 154–159.
- Fullerton DG, Bruce N and Gordon SB. Indoor air pollution from biomass fuel smoke is a major health concern in the developing world. *Trans R Soc Trop Med Hyg* 2008; 102: 843–851.
- Kena T, Abebe Y and Alem M. Effects of indoor air pollution by biomass fuels on respiratory functions of women in Gondar, north west Ethiopia. *Int J Pharm Ind Res* 2013; 3: 232–242.
- Coselman KE, Navas-Acien A and Kaufman JD. Environmental factors in cardiovascular disease. *Nat Rev Cardiol* 2015; 12: 627–642.
- Muralidharan V, Sussan TE, Limaye S, Koehler K, Williams DL, Rule AM, Juvekar S, Breyse PN, Salvi S and Biswal S. Field testing of alternative cookstove performance in a rural setting of western India. *Int J Environ Res Public Health* 2015; 12: 1773–1787.
- Umoh VA and Peters E. The relationship between lung function and indoor air pollution among rural women in the Niger Delta region of Nigeria. *Lung India* 2014; 31: 110–115.
- Guarnieri MJ, Diaz JV, Basu C, Diaz A, Pope D, Smith KR, Smith-Sivertsen T, Bruce N, Solomon C, McCracken J, Balmes JR. Effects of woodsmoke exposure on airway inflammation in rural Guatemalan women. *PLoS One* 2014; 9: 1–9.
- Adeniji BA, Ana Gree Adedokun BO and Ige OI. Exposure to emissions from kerosene cooking stoves and the pulmonary health status of women in Olorunda Community, Ibadan, Nigeria. *J Environ Prot (Irvine, Calif)* 2015; 6: 435–445.
- Omaka ON, Nwabue FI, Itumoh EJ and Okeke GN. Gas emissions and metallic contents of commonly used fuelwood in Nigeria. *Environ Pollut* 2013; 2: 100–106.
- Kilabuko JH and Nakai S. Effects of cooking fuels on acute respiratory infections in children in Tanzania. *Int J Environ Res Public Health* 2007; 4: 283–288.
- Pope D, Diaz E, Smith-Sivertsen T, Lie RT, Bakke P, Balmes JR, Smith KR and Bruce NG. Exposure to household air pollution from wood combustion and association with respiratory symptoms and lung function in nonsmoking women: results from the randomized exposure study of pollution indoors and respiratory effects trial, Guatemala. *Environ Health Perspect* 2015; 123: 285–292.
- Faris K. Survey of indoor air pollution problems in rural communities of Jimma, Southwest Ethiopia. *Ethiop J Health Sci* 2002; 12: 1–14.
- Roba C, Stefanie H, Torok Z, Kovacs M, Rosu C and Ozunu A. Determination of volatile organic compounds and particulate matter levels in an urban area from Romania. *Environ Eng Manag J* 2014; 13: 2261–2268.
- Tahri M, Bounakhla M, Zghaid M, Noack Y, Benyaich F and Benchrif A. Evaluation of airborne particulate matter pollution in Kenitra City, Morocco. *Ambi-Água – Inter J Appl Sci* 2013; 8: 38–47.
- Ojiodu CC. Ambient volatile organic compounds (VOCs) pollution in Isolo industrial area of Lagos State, Southwestern Nigeria. *EJESM* 2013; 6: 688–697.
- Do DH, Van Langenhove H, Walgraeve C, Hayleeyesus SF, De Wispelaere P, Dewulf J and Demeestere K. Volatile organic compounds in an urban environment: a comparison among Belgium, Vietnam and Ethiopia. *Int J Environ an Ch* 2013; 93: 298–314.
- Int Panis L, de Geus B, Vandenbulcke G, Willems H, Degraeuwe B, Bleux N, Mishra V, Thomas I and Meeusen R. Exposure to particulate matter in traffic: a comparison of cyclists and car passengers. *Atmos Environ* 2010; 44: 2263–2270.
- Ahmed MJ, Ali M, Hossain M, Siraj S and Ahsan MA. Determination of trace metals in air of Chittagong city-Bangladesh. *Eur J Chem* 2012; 3: 416–420.
- Aryal R, Kim A, Lee B-K, Kamruzzaman M and Beecham S. Characteristics of atmospheric particulate matter and metals in industrial sites in Korea. *Environ Pollut* 2013; 2: 10–21.
- Krzemińska-Flowers M, Bem H and Górecka H. Trace metals concentration in size-fractioned urban air particulate matter in Lodz, Poland. I. Seasonal and site fluctuations. *Polish J Environ Stud* 2006; 15: 759–767.
- Uno UA, Ekpo BO, Etuk VE, Etuk HS and Ibok UJ. Comparative study of levels of trace metals in airborne particulates in some cities of the Niger delta region of Nigeria. *Environ Pollut* 2013; 2: 110–121.
- Zabiegała B, Partyka M, Zygmunt B and Namiesnik J. Determination of volatile organic compounds in indoor

- air in the Gdansk area using permeation passive samplers. *Indoor Built Environ* 2009; 18: 492–504.
26. US EPA. Integrated review plan for the national ambient air quality standards for particulate matter, [www3.epa.gov/ttn/naaqs/standards/pm/data/2008\\_03\\_final\\_integrated\\_review\\_plan.pdf](http://www3.epa.gov/ttn/naaqs/standards/pm/data/2008_03_final_integrated_review_plan.pdf) (accessed 5 January 2018).
  27. WHO. Exposure to air pollution: a major public health concern, [www.who.int/ipcs/features/air\\_pollution.pdf](http://www.who.int/ipcs/features/air_pollution.pdf) (accessed 5 January 2018).
  28. Etyemezian V, Tesfaye M, Yimer A, Chow J, Mesfin D, Nega T, Nikolich G, Watson J and Wondmagegn M. Results from a pilot-scale air quality study in Addis Ababa, Ethiopia. *Atmos Environ* 2005; 39: 7849–7860.
  29. Graham M. *A mixed methods approach to assessing indoor air pollution among women in Addis Ababa, Ethiopia*. Master's Thesis, Addis Ababa University, Ethiopia, 2011.
  30. Hailu AT, Chaubey AK and Hibstie AY. Investigation of traffic air pollution in Addis Ababa city around selected bus stations using instrumental neutron activation technique. *Int J Basic Appl Sci* 2013; 1: 613–620.
  31. Kumie A, Emmelin A, Wahlberg S, Berhane Y, Ali A, Mekonen E, Worku A and Brandstrom D. Sources of variation for indoor nitrogen dioxide in rural residences of Ethiopia. *Environ Health* 2009; 8: 1–11.
  32. Gebre G, Feleke Z and Sahle-Demissie E. Mass concentrations and elemental composition of urban atmospheric aerosols in Addis Ababa, Ethiopia. *Bull Chem Soc Ethiop* 2010; 24: 361–373.
  33. Devi JJ, Gupta T, Tripathi SN and Ujwal KK. Assessment of personal exposure to inhalable indoor and outdoor particulate matter for student residents of an academic campus (IIT-Kanpur). *Inhal Toxicol* 2009; 21: 1208–1222.
  34. Cattaneo A, Taronna M, Consonni D, Angius S, Costamagna P and Cavallo DM. Personal exposure of traffic police officers to particulate matter, carbon monoxide, and benzene in the city of Milan, Italy. *J Occup Environ Hyg* 2010; 7: 342–351.
  35. Rabinovitch N, Adams CD, Strand M, Koehler K and Volckens J. Within-microenvironment exposure to particulate matter and health effects in children with asthma: a pilot study utilizing real-time personal monitoring with GPS interface. *Environ Health* 2016; 15: 1–10.
  36. Son JY and Bell ML. The relationships between short-term exposure to particulate matter and mortality in Korea: Impact of particulate matter exposure metrics for sub-daily exposures. *Environ Res Lett* 2013; 8: 1–8.
  37. Gulliver J and Briggs DJ. Personal exposure to particulate air pollution in transport microenvironments. *Atmos Environ* 2004; 38: 1–8.
  38. Wangchuk T, Mazaheri M, Clifford S, Dudzinska MR, He C, Buonanno G and Morawska L. Children's personal exposure to air pollution in rural villages in Bhutan. *Environ Res* 2015; 140: 691–698.
  39. Danielle. Misir and Shiro Wat Recipes, <http://asoufulfulpetite.com/african-cuisine/misir-and-shiro-wat-ethiopian-101/> (accessed 2 December 2017).
  40. Sanbata H, Asfaw A and Kumie A. Indoor air pollution in slum neighbourhoods of Addis Ababa, Ethiopia. *Atmos Environ* 2014; 89: 230–234.
  41. Martinez MN and Bartholomew MJ. What does it “mean”? A review of interpreting and calculating different types of means and standard deviations. *Pharmaceutics* 2017; 9: 1–23.
  42. Cao C, Gao J, Wu L, Ding X and Zhang X. Ventilation improvement for reducing individual exposure to cooking-generated particles in Chinese residential kitchen. *Indoor Built Environ* 2016; 26: 226–237.
  43. Balakrishnan K, Parikh J, Sankar S, Padmavathi R, Srividya K, Venugopal V, Prasad S and Pandey VL. Daily average exposures to respirable particulate matter from combustion of biomass fuels in rural households of Southern India. *Environ Health Perspect* 2002; 110: 1069–1075.
  44. McDonald JD, Zielinska B, Fujita EM, Sagebiel JC, Chow JC and Watson JG. Fine particle and gaseous emission rates from residential wood combustion. *Environ Sci Technol* 2000; 34: 2080–2091.
  45. Bo M, Salizzoni P, Clerico M and Buccolieri R. Assessment of indoor-outdoor particulate matter air pollution: a review. *Atmosphere* 2017; 8: 1–18.
  46. Zhao WC, Cheng JP, Yu ZY, Tang QL, Cheng F, Yin YW and Wang WH. Levels, seasonal variations, and health risks assessment of ambient air pollutants in the residential areas. *Int J Environ Sci Technol* 2013; 10: 487–494.
  47. Ezzati M, Saleh H and Kammen DM. The contributions of emissions and spatial microenvironments to exposure to indoor air pollution from biomass combustion in Kenya. *Environ Health Perspect* 2000; 108: 833–839.
  48. Yeung LL and To WM. Size distributions of the aerosols emitted from commercial cooking processes. *Indoor Built Environ* 2008; 17: 220–229.
  49. Jin Y, Zhou Z, He G, Wei H, Liu J, Liu F, Tang N, Ying B, Liu Y, Hu G, Wang H, Balakrishna K, Watson K, Baris E, Ezzati M. Geographical, spatial, and temporal distributions of multiple indoor air pollutants in four Chinese provinces. *Environ Sci Technol* 2005; 39: 9431–9439.
  50. Saksena S, Prasad RK and Shankar VR. Daily exposure to air pollutants in indoor, outdoor and in-vehicle microenvironments: a pilot study in Delhi. *Indoor Built Environ* 2007; 16: 39–46.
  51. Balakrishnan K, Sambandam S, Ramaswamy P, Mehta S and Smith KR. Exposure assessment for respirable particulates associated with household fuel use in rural districts of Andhra Pradesh, India. *JEAE* 2004; 14: 14–25.
  52. Dasgupta S, Huq M, Khaliquzzaman M, Pandey K and Wheeler D. Indoor air quality for poor families: new evidence from Bangladesh. *Indoor Air* 2006; 16: 426–444.
  53. Bautista LE, Correa A, Baumgartner J, Breyse P and Matanoski GM. Indoor charcoal smoke and acute respiratory infections in young children in the Dominican Republic. *Am J Epidemiol* 2009; 169: 572–580.
  54. Kushwaha R, Lal H, Srivastava A and Jain VK. Human exposure to particulate matter and their risk assessment over Delhi, India. *Natl Acad Sci Lett* 2012; 35: 497–504.

55. Matookane M and Diab R. Health risk assessment for sulfur dioxide pollution in South Durban, South Africa. *Arch Environ Health* 2003; 58: 763–770.
56. Morakinyo OM, Adebowale AS, Mokgobu MI and Mukhola S. Health risk of inhalation exposure to sub-10  $\mu\text{m}$  particulate matter and gaseous pollutants in an urban-industrial area in South Africa: an ecological study. *BMJ Open* 2017; 7: 1–9.
57. US EPA. Risk assessment guidance for superfund. volume I: human health evaluation manual (Part A), [www.epa.gov/sites/production/files/2015-09/documents/rags\\_a.pdf](http://www.epa.gov/sites/production/files/2015-09/documents/rags_a.pdf) (accessed 5 January 2018).
58. Walpole SC, Prieto-Merino D, Edwards P, Cleland J, Stevens G and Roberts I. The weight of nations: an estimation of adult human biomass. *BMC Public Health* 2012; 12: 1–6.
59. Ncube E and Phiri B. Concentrations of heavy metals in Eucalyptus and Pinus wood sawdust and smoke, Copperbelt province, Zambia. *Maderas, Cienc Tecnol* 2015; 17: 585–596.
60. Schmidl C, Bauer H, Dattler A, Hitznerberger R, Weissenboeck G, Marr IL and Puxbaum H. Chemical characterisation of particle emissions from burning leaves. *Atmos Environ* 2008; 42: 9070–9079.
61. Urbanski SP, Hao WM and Baker SC. 4 chemical composition of wildland fire emissions in: developments in environmental science Vol. 8. In: Bytnerowicz A, Arbaugh M, Riebau A and Andersen C (eds) Elsevier: Online Publication, 2008; pp. 79–107. DOI: 10.1016/S1474-8177(08)00004-1.